## ORIGINAL ARTICLE Microbial mutualism dynamics governed by dose-dependent toxicity of cross-fed nutrients

Breah LaSarre, Alexandra L McCully, Jay T Lennon and James B McKinlay Department of Biology, Indiana University, Bloomington, IN, USA

Microbial interactions, including mutualistic nutrient exchange (cross-feeding), underpin the flow of energy and materials in all ecosystems. Metabolic exchanges are difficult to assess within natural systems. As such, the impact of exchange levels on ecosystem dynamics and function remains unclear. To assess how cross-feeding levels govern mutualism behavior, we developed a bacterial coculture amenable to both modeling and experimental manipulation. In this coculture, which resembles an anaerobic food web, fermentative Escherichia coli and photoheterotrophic Rhodopseudomonas palustris obligately cross-feed carbon (organic acids) and nitrogen (ammonium). This reciprocal exchange enforced immediate stable coexistence and coupled species growth. Genetic engineering of R. palustris to increase ammonium cross-feeding elicited increased reciprocal organic acid production from E. coli, resulting in culture acidification. Consequently, organic acid function shifted from that of a nutrient to an inhibitor, ultimately biasing species ratios and decreasing carbon transformation efficiency by the community; nonetheless, stable coexistence persisted at a new equilibrium. Thus, disrupting the symmetry of nutrient exchange can amplify alternative roles of an exchanged resource and thereby alter community function. These results have implications for our understanding of mutualistic interactions and the use of microbial consortia as biotechnology. The ISME Journal (2017) 11, 337-348; doi:10.1038/ismej.2016.141; published online 29 November 2016

### Introduction

Ecosystems comprise a large, variable number of species that interact with both each other and their environment. Metabolic interactions between species, such as detoxification of metabolic waste and cross-feeding, shape microbial communities and regulate ecosystem processes (Schink, 2002: Fuhrman, 2009; Morris et al., 2013). Study of these interactions can be encumbered by the stochasticity and complexity of natural systems. As an alternative, synthetic microbial communities (cocultures) have been developed in which two or more species are cultivated together under laboratory conditions. Cocultures preserve core aspects of natural systems while offering greater practical experimental control (Shou et al., 2007; Hillesland and Stahl, 2010; Summers et al., 2010; Harcombe, 2010; Momeni et al., 2011; Hom and Murray, 2014; Mee et al., 2014). They are also more amenable to modeling than are natural systems and facilitate the development, experimental testing and refining of models for predicting community behavior (Zomorrodi and Segrè, 2015; Johns et al., 2016; Lindemann et al., 2016;

Widder *et al.*, 2016). As such, cocultures are very useful for defining the principles that underlie the stability and performance of microbial communities.

One factor that can promote cross-feeding in natural ecosystems is nutrient limitation, as crossfeeding can be a means by which to acquire scarce key nutrients (Shou et al., 2007; Klitgord and Segrè, 2010; Hom and Murray, 2014; Zelezniak et al., 2015). Limitation of bacterial growth is most commonly considered to result from nutrient deficiency; however, growth can also be influenced by diverse compounds that are inhibitory when abundant but serve as nutrients at lower concentrations (Abbott, 1973; Kunz et al., 1992; Alvarez et al., 2009; Barnhill et al., 2010). Organic acids, which are key intermediates in global carbon cycles (McInerney et al., 2009), exemplify this paradox in both natural and applied settings (Lee et al., 1976; Louis et al., 2007; Sousa et al., 2009; Huang et al., 2011). There is limited knowledge of how potentially toxic, crossfed nutrients such as organic acids impact mutualism dynamics and stability. However, given the ubiquity of organic acid exchange, their impact may be profound, as the cross-feeding levels compatible with cooperation would be constrained between thresholds of inadequate exchange and toxicity.

Cocultures pairing fermentative bacteria with purple photoheterotrophic bacteria have been of interest for over 30 years (Odom and Wall, 1983;

Correspondence: JB McKinlay, Department of Biology, Indiana University, 1001 E 3rd Street, Jordan Hall, Bloomington, IN 47405, USA.

E-mail: jmckinla@indiana.edu

Received 8 May 2016; revised 25 August 2016; accepted 7 September 2016; published online 29 November 2016

Fang et al., 2006; Ding et al., 2009; Sun et al., 2010; Jiao et al., 2012), primarily as a consolidated bioprocess for converting plant-derived sugars into  $H_2$ . However, this combination of bacteria also resembles a natural anaerobic food web; fermentative bacteria consume plant-derived sugars and excrete products that serve as a carbon source for photoheterotrophs (Figure 1a). Despite decades of research, coculture instability has limited the utility of these cocultures for both practical applications, such as  $H_2$  production, and fundamental studies into anaerobic microbial cross-feeding.

Here, we use a bacterial coculture based on fermentative Escherichia coli and purple phototrophic *Rhodopseudomonas palustris* to assess how the dynamics and function of a mutualism are influenced by the degree of cross-feeding. In this coculture, E. coli and R. palustris stably coexist due to the obligate bidirectional exchange of two essential nutrients, carbon and nitrogen (Figure 1b). Using a combination of experimental and modeling approaches, we show that enhanced cross-feeding can disproportionately inhibit members of a mutualism by amplifying toxic attributes of an exchanged nutrient, with detrimental consequences for the community transformation of carbon into biomass and CO<sub>2</sub>. Nonetheless, obligate cross-feeding can uphold stable coexistence amid detrimental exchange levels.

### Materials and methods

#### Strains, plasmids and growth conditions

Strains, plasmids and primers are listed in Supplementary Table 1. *E. coli* and *R. palustris* were cultivated on Luria-Burtani (LB) agar or defined mineral (PM) (Kim and Harwood, 1991) agar with 10 mm succinate, respectively. These media were also used to selectively plate for each species after coculturing, except that  $(NH_4)_2SO_4$  was excluded from PM plates to select for R. palustris. E. coli plates were incubated aerobically at 30 °C in the dark, whereas R. palustris plates were incubated anaerobically at 30 °C in a jar with a GasPak sachet (BD, Franklin Lakes, NJ, USA) in front of a 60 W light bulb. Cloning was performed using E. coli NEB10β (New England Biolabs, Ipswich, MA, USA) or E. coli XL10Gold (Stratagene, La Jolla, CA, USA). When necessary, gentamicin (Gm) was included at 15 µg ml<sup>-1</sup> for  $\tilde{E}$ .  $\tilde{coli}$  or 100 µg ml<sup>-1</sup> for R. palustris.

Cultures were grown in 10 ml of defined coculture medium (MDC) in 27-ml anaerobic test tubes except for time-course analyses, which used 60 ml MDC in 160-ml serum vials. MDC contained Na<sub>2</sub>HPO<sub>4</sub> (42.5 mM), KH<sub>2</sub>PO<sub>4</sub> (22 mM), Na<sub>2</sub>S<sub>2</sub>O<sub>3</sub> (0.1 mM), *p*-aminobenzoic acid (14.6  $\mu$ M) and trace elements solution (0.1% v/v; Kremer *et al.*, 2015). MDC was made anaerobic by bubbling with 100% N<sub>2</sub> and then sealing with rubber stoppers and aluminum crimps



**Figure 1** Imposing bidirectional cross-feeding changes coculture dynamics. (**a**, **b**) Cross-feeding under conditions compelling unidirectional transfer of carbon to *R. palustris* and nitrogen to *E. coli* (**b**). The latter requires a NH<sub>4</sub><sup>+</sup>-excreting *R. palustris* strain. Filled bubbles indicate externally added compounds. (**c**) Representative growth curves for NH<sub>4</sub><sup>+</sup>-supplied cocultures (15 mM NH<sub>4</sub><sup>+</sup> and 100% N<sub>2</sub> headspace; open circles) and N<sub>2</sub>-supplied cocultures (100% N<sub>2</sub> headspace; closed circles) containing either parental *R. palustris* (CGA4004; black), which does not share NH<sub>4</sub><sup>+</sup>, or *R. palustris* Nx (CGA4005; green), which shares NH<sub>4</sub><sup>+</sup>. Log transformed growth curves are in Supplementary Figure 3. All cocultures were inoculated at a 1:1 species ratio and contained 25 mM glucose. (**d**) Final *E. coli* percentage in NH<sub>4</sub><sup>+</sup>-supplied (solid bars) or N<sub>2</sub>-supplied (hatched bars) cocultures with *R. palustris* NX inoculated with different starting species ratios. A starting value of 1 corresponds to ~ 2.5 × 10<sup>8</sup> CFU ml<sup>-1</sup>. ND, not determined *as E. coli* dominated NH<sub>4</sub><sup>+</sup>-supplied cocultures even when starting ratios did not favor *E. coli*. Ec, *E. coli*; Rp, *R. palustris*. Error bars, s.e.m., *n*=3.

 $(100\% N_2$  headspace). After autoclaving, MDC was supplemented with cation solution (1% v/v; 100 mM MgŜO₄ and 10 mм CaCl₂) and glucose (25 mм) if used for cocultures. MDC also received either NH<sub>4</sub>Cl (15 mm, conditions which we refer to as  $NH_{4}^{+}$ supplied cocultures) or NaCl (15 mm, conditions which we refer to as N<sub>2</sub>-supplied cocultures). For MOPS-supplemented cocultures, MOPS buffer (100 mM final concentration), pH 7 was added to MDC before autoclaving. For spent E. coli supernatant experiments, E. coli cultures were grown to stationary phase in MDC supplemented with cation solution, 25 mM glucose and 15 mM NH<sub>4</sub>Cl. Following removal of E. coli cells by filtration, spent supernatants were either supplemented with 100 mM MOPS buffer, pH 7 (+MOPS) or an equivalent volume of MDC (-MOPS). Spent supernatants were then injected into sterile anoxic tubes and flushed with 100% N<sub>2</sub> before inoculation with R. palustris cells.

Final cell densities in starter monocultures were controlled using carbon limitation (MDC with 3 mM acetate) for *R. palustris* or nitrogen limitation (MDC with 1% v/v cation solution, 25 mM glucose and 2 mM NH<sub>4</sub>Cl) for *E. coli*. For non-growing *E. coli* suspensions, nitrogen-limited *E. coli* cultures were grown to stationary phase. Cells were then collected, washed and resuspended in 1 ml MDC, injected into fresh MDC with glucose but lacking NH<sup>‡</sup>, and then flushed with 100% N<sub>2</sub>. All cultures and cell suspensions were incubated at 30 °C laying flat under a 60 W incandescent bulb with shaking at 150 r.p.m.

#### *Generation of* R. palustris *mutants*

*R. palustris* mutants were derived from wild-type CGA009 (Larimer *et al.*, 2004). The vector used to introduce the *nifA*<sup>\*</sup> mutation was described previously (McKinlay and Harwood, 2010). Vectors used to delete other genes were generated by PCR amplifying regions flanking the gene to be deleted using primers in Supplementary Table 1. Product pairs were fused by overlap extension PCR and subsequently cloned into pJQ200SK. Vectors were introduced into *R. palustris* by conjugation with *E. coli* S17-1 (Rey *et al.*, 2006) or by electroporation (Pelletier *et al.*, 2008). Mutants were generated using sequential selection and screening as described (Rey *et al.*, 2006). Mutant genotypes were confirmed by PCR and sequencing.

#### Generation of amtB2 complementation vector

A 2.2 kb region encompassing *glnK2*, *amtB2* and their shared promoter was amplified using primers BL457 and BL498 (Supplementary Table 1) and cloned into pBBR1MCS-5 to generate pBBRGlnkAmtB2. To prevent GnlK2 over-expression, the *gnlK2* start codon in pBBRGlnkAmtB2 was mutated to GGG by single-primer site-directed mutagenesis using primer BL499,

resulting in pBBRAmtB2. The desired mutation was confirmed by sequencing.

#### Coculture inoculation

Cocultures were inoculated using either specific densities, single colonies or subculturing as indicated in figure legends. Cocultures inoculated with specific densities were generated by first growing monocultures in MDC with limiting nitrogen or carbon for E. coli or R. palustris, respectively. Monocultures were then diluted with MDC until the appropriate cell densities were achieved, as estimated from experimentally determined speciesspecific  $OD_{660}/CF\bar{U}$  per ml standard curves. For cocultures started from colonies, single colonies of each species were combined in 100 µl MDC for use as the inoculum. For subculturing, stationary phase cocultures were subcultured 1:100 into fresh medium. Growth and metabolic trends from cocultures inoculated by each of these methods were comparable.

#### Analytical procedures

Cell density was assaved by optical density at  $660 \text{ nm} (OD_{660})$  using a Genesys 20 visible spectrophotometer (Thermo-Fisher, Waltham, MA, USA). Most growth curve readings were taken in culture tubes without sampling. Specific growth rates were determined using tube OD measurements between 0.1 and 1.0  $OD_{660}$  where there is linear correlation between cell density and OD<sub>660</sub>. Final cell densities and all serum vial readings were taken in cuvettes and samples were diluted as necessary to achieve values below 1  $OD_{660}$ . H<sub>2</sub> was quantified using a Shimadzu (Kyoto, Japan) gas chromatograph as described (Huang *et al.*, 2010). Glucose, organic acids, and ethanol were quantified using a Shimadzu high-performance liquid chromatograph (HPLC) as described (McKinlay et al., 2005). NH<sub>4</sub><sup>+</sup> was quantified using an indophenol colorimetric assay as follows. Cultures were sampled 15-30 h into stationary phase. Supernatant (550 µl) was mixed with 50 µl 1 м NaOH, 100 µl phenol nitroprusside (Sigma, St Louis, MO, USA), and 100 µl alkaline hypochlorite (Sigma), followed by incubation at room temperature for 15 min. The  $A_{630nm}$  of each sample was converted into  $\mu M NH_4^+$  using a standard curve and then normalized to cell density  $(OD_{660})$ .

#### Determination of species ratios and cell densities

All final species ratios were determined using fluorescence *in situ* hybridization and microscopy (FISH). Cells were fixed, washed, stored and dehydrated for hybridization as described (Oda *et al.*, 2000). Dried cell pellets were resuspended in 50 µl hybridization buffer (Oda *et al.*, 2000) incubated at 37 °C for 30 min, and then hybridized for 2.5 h at 60 °C with an *E. coli* specific probe (ALM-Ec3,

4.5 ng µl<sup>-1</sup>, ATTO550-5'-TGCTCTCGCGAGGTCGCT TC-3'; ATTO-TEC, Siegen, Germany). Probe specificity was empirically confirmed using E. coli and R. palustris monocultures. Hybridized samples were washed three times with hybridization buffer and then resuspended in 5 µM SYTO9 (Life Technologies, Carlsbad, CA, USA) to stain both species. At least 2000 cells were visualized per sample by epifluorescence microscopy (Ti-E; Nikon, Tokyo, Japan) within 72 h of hybridization. Cell counting was partially automated using Microbe J (Ducret et al., 2016). E. coli frequency was determined from the number of ATTO550-positive cells per total number of Syto9positive cells counted in the sample. Species ratios determined using FISH were periodically confirmed by CFU enumeration for corresponding live cocultures. Intermediate species ratios as well as monoculture and coculture cell densities were determined by selective plating and CFU enumeration.

### Mathematical modeling

Ecological models A and B were based on Monod models used to describe microbial mutualism (Meyer et al., 1975; Lee et al., 1976) but were modified as follows: (i) the dilution parameter was omitted to describe batch cultures rather than continuous cultures; (ii) growth-independent fermentation para-meters were included ('r', Supplementary Table 2); (iii) a parameter to modulate *E. coli* product excretion between growth-dependent and growthindependent levels was included ('ng' used in the term 'ng/( $\mu$ +ng)', Supplementary Table 2); and (iv) model A also includes parameters to describe the growth-inhibiting pH effects of organic acids on both species ('b' used in the term  $b/(b+10^{(f+C)})$ ' Supplementary Table 2); this inhibition was omitted in model B. Growth inhibition by ethanol was not included as ethanol levels never reach those we determined to inhibit growth in monoculture. The models were developed in R studio and are available for download at https://github.com/McKinlab/Cocul ture-Mutualism. Default parameter values and descriptions are in Supplementary Table 2. The differential equations used in the models are as follows:

$$\begin{split} E. \ coli \ \text{growth rate}; \mu_{\text{Ec}} &= \mu_{\text{EcMAX}} \\ & \cdot \left[ G/(k_{\text{G}} + G) \right] \cdot \left[ A/(k_{\text{A}} + A) \right] \\ & \cdot \left[ b_{\text{Ec}} / \left( b_{\text{Ec}} + \mathbf{10}^{(\text{f+C})} \right) \right] \end{split}$$
(1)

$$\begin{aligned} R. \ palustris \ \text{growth rate}; \mu_{\text{Rp}} &= \mu_{\text{RpMAX}} \cdot [C/(k_{\text{C}} + C)] \\ & \cdot [N/(k_{\text{N}} + N)] \\ & \cdot \left[ b_{\text{Rp}} / \left( b_{\text{Rp}} + \mathbf{10}^{(\text{f+C})} \right) \right] \end{aligned}$$

$$(2)$$

Equations (3–12) describe the changes in cell densities and extracellular compounds through time.

donortioo una o

The ISME Journal

Only net accumulation of formate, ethanol,  $CO_2$  and  $H_2$  are described in accordance with observed trends. Numerical constants in product excretion equations are used to account for the moles of product carbon relative to glucose. All R and r parameters are expressed in stoichiometric terms of glucose consumed except for  $R_A$ , which is the amount of  $NH_4^+$  produced per R. palustris cell (Supplementary Table S2).

 $E. \ coli; dEc/dt = \mu_{Ec} \cdot Ec \tag{3}$ 

$$R. \ palustris; \ dRp/dt = \mu_{Rp} \cdot Rp$$
(4)

$$\begin{aligned} \text{Glucose; } \text{d}G/\text{d}t &= -\mu_{\text{Ec}} \cdot \text{Ec}/Y_{\text{G}} - \mu_{\text{Ec}} \cdot \text{Ec} \\ & \cdot (R_{\text{c}} + R_{\text{f}} + R_{\text{e}} + R_{\text{CO}_2}) - \text{Ec} \\ & \cdot (r_{\text{C}} + r_{\text{f}} + r_{\text{e}} + r_{\text{CO}_2}) \\ & \cdot [G/(k_{\text{G}} + G)] \cdot [\text{ng}/(\mu_{\text{Ec}} + \text{ng})] \end{aligned} \tag{5}$$

$$N_2; dN/dt = -\mu_{Rp} \cdot Rp/Y_N - \mu_{Rp} \cdot Rp \cdot 0.5 \cdot R_A$$
(6)

Consumable organic acids; dC/dt

$$= \mu_{\rm Ec} \cdot {\rm Ec} \cdot R_{\rm C} \cdot 2 + {\rm Ec} \cdot r_{\rm C} \cdot 2 \cdot [G/(k_{\rm G} + G)]$$
  
 
$$\cdot [{\rm ng}/(\mu_{\rm Ec} + {\rm ng})] - \mu_{\rm Rp} \cdot {\rm Rp}/Y_{\rm C}$$
(7)

$$\begin{split} \text{Formate; } \mathrm{d}f/\mathrm{d}t &= \mu_{\mathrm{Ec}} \cdot \mathrm{Ec} \cdot R_{\mathrm{f}} \cdot \mathbf{6} + \mathrm{Ec} \cdot r_{\mathrm{f}} \cdot \mathbf{6} \\ & \cdot \left[ G/(k_{\mathrm{G}}+G) \right] \cdot \left[ \mathrm{n}g/(\mu_{\mathrm{Ec}}+\mathrm{n}g) \right] \quad (\mathbf{8}) \end{split}$$

Ethanol; 
$$de/dt = \mu_{\text{Ec}} \cdot \text{Ec} \cdot R_{\text{e}} \cdot 3 + \text{Ec} \cdot r_{\text{e}} \cdot 3$$
  
  $\cdot [G/(k_{\text{G}} + G)] \cdot [\text{ng}/(\mu_{\text{Ec}} + \text{ng})]$  (9)

$$CO_2; dCO_2/dt = \mu_{Ec} \cdot Ec \cdot R_{CO_2} \cdot 6 + Ec \cdot r_{CO_2} \cdot 6$$
$$\cdot [G/(k_G + G)] \cdot [ng/(\mu_{Ec} + ng)] \quad (10)$$

$$\mathrm{NH_4}^+; \mathrm{d}A/\mathrm{d}t = \mu_{\mathrm{Rp}} \cdot \mathrm{Rp} \cdot R_{\mathrm{A}} - \left[(\mu_{\mathrm{Ec}} \cdot \mathrm{Ec})/Y_{\mathrm{A}}\right] \qquad (11)$$

 $\mathrm{H}_{2};\mathrm{d}H/\mathrm{d}t = \mu_{\mathrm{Rp}}\cdot\mathrm{Rp}\cdot\mathrm{Rp}\cdot\mathrm{R}_{\mathrm{HRp}} + \mu_{\mathrm{Ec}}\cdot\mathrm{Ec}\cdot\mathrm{R}_{\mathrm{HEc}}$ 

$$+ \operatorname{Ec} \cdot r_{\mathrm{H}} \cdot [G/(k_{\mathrm{G}} + G)] \\ \cdot [\operatorname{ng}/(\mu_{\mathrm{Ec}} + \operatorname{ng})]$$
(12)

where,

 $\mu$  is the specific growth rate of the indicated species (h<sup>-1</sup>).

 $\mu_{MAX}$  is the maximum specific growth rate of the indicated species (h<sup>-1</sup>), based on growth rates measured in monocultures in MDC with acetate and 100% N<sub>2</sub> for *R. palustris* or glucose and NH<sub>4</sub>Cl for *E. coli*.

G, A, C, N, f, e and  $CO_2$  are the concentrations (mM) of glucose, NH<sub>4</sub><sup>+</sup>, consumable organic acids, N<sub>2</sub> (assumed to be fully dissolved), formate, ethanol and  $CO_2$ , respectively. Consumable organic acids are those that *R. palustris* consumes, namely, lactate (3 carbons), acetate (2 carbons) and succinate (4 carbons). All consumable organic acids were simulated to have three carbons for convenience.

k is the half saturation constant for the indicated substrate (m<sub>M</sub>).

Ec and Rp are the cell densities (cells  $ml^{-1}$ ) of *E. coli* and *R. palustris*, respectively.

ng is used to modulate the transition between values used to describe product excretion by *E. coli* in the presence and absence of growth  $(h^{-1})$ .

b is used to modulate the extent to which an indicated species resists growth-inhibiting effects of organic acids (mm).

Y is the *E. coli* or *R. palustris* cell yield from the indicated substrate (cells  $\mu$ mol<sup>-1</sup> glucose). Values were determined for each species in MDC with the indicated substrate as the limiting nutrient.

*R* is the fraction of glucose converted into the indicated compound per *E. coli* cell during growth (µmol cell<sup>-1</sup>), except for  $R_A$ . Values were based on product yields measured in monocultures in MDC with acetate and 100% N<sub>2</sub> for *R. palustris* or glucose and NH<sub>4</sub>Cl for *E. coli*.

 $R_{\rm A}$  is the ratio of NH<sup>+</sup><sub>4</sub> produced per *R. palustris* cell during growth (µmol cell<sup>-1</sup>). The default value was based on that which led to simulations resembling observed trends.

r is the growth-independent rate of glucose converted into the indicated compound by *E. coli* (µmol cell<sup>-1</sup>h<sup>-1</sup>). Specific production rates for nongrowing *E. coli* (r) were determined in nitrogen-free cell suspensions by linear regression between days 1–7 (Supplementary Figure 1). *E. coli* product excretion rates were increased twofold from those measured in cell suspensions as we assumed that higher formate yield observed in coculture was true of other products as well. Ethanol and H<sub>2</sub> production rates were unchanged from monoculture observations to match coculture observations.

### Results

# Obligate mutualism enforces stable coexistence in coculture by means of coupled growth

Cocultures pairing fermentative and purple photoheterotrophic bacteria resemble natural anaerobic foodwebs where organic acid fermentation products serve as a carbon source for the phototroph (Figure 1a). Over the last 30 years, traditional growth conditions for these cocultures promoted interactions in which phototrophs relied on fermentative bacteria but provided no reciprocal benefit (Figure 1a; Odom and Wall, 1983; Fang *et al.*, 2006; Ding *et al.*, 2009; Sun *et al.*, 2010). Consequently, these systems commonly suffered from instability and variable product yields (Fang *et al.*, 2006; Ding *et al.*, 2009; Sun *et al.*, 2010).

We overcame this long-standing stability issue by exploiting two metabolic features of the phototrophic purple nonsulfur bacterium, *R. palustris*. First, *R. palustris* can consume fermentation products but not glucose (Larimer *et al.*, 2004). This means that *R. palustris* cannot compete against *E. coli* for glucose in coculture but is instead reliant on excreted *E. coli* fermentation products. Second, *R. palustris* is a diazotroph, and thus can use the enzyme, nitrogenase, to convert N<sub>2</sub> into bioavailable ammonium  $(NH_4^+)$  in a process known as  $N_2$  fixation. We found that a NifA<sup>\*</sup> mutant exhibiting constitutive nitrogenase activity (McKinlay and Harwood, 2010) excreted measurable  $NH_4^+$  when fixing  $N_2$ (Supplementary Figure 2). This R. palustris strain, hereon referred to as Nx, has a NifA\* mutation to allow for NH<sub>4</sub><sup>+</sup> excretion and  $\Delta hupS$  and  $\Delta rpa2750$ mutations to prevent  $H_2$  uptake (Rey *et al.*, 2006) and cell aggregation (Fritts et al., in preparation), respectively. In contrast to R. palustris, E. coli cannot fix  $N_2$ . Thus, we reasoned that  $NH_4^+$  excreted by *R. palustris* could cross-feed *E. coli* and establish an obligate mutualism (that is, syntrophy; Figure 1b). Indeed, we observed robust growth when we paired E. coli with R. palustris Nx in a defined medium (MDC) supplemented with glucose as the sole carbon source and with  $N_2$  gas as the sole nitrogen source  $(N_2$ -supplied coculture). In contrast, growth was negligible when E. coli was paired with the parent R. palustris strain that does not share  $NH_4^+$ (Figure 1c). Our results demonstrate that under these nutrient conditions bidirectional cross-feeding is necessary for cooperative coculture growth; R. *palustris* relies on *E. coli* for essential carbon in the form of fermentation products, whereas *E. coli* relies on *R. palustris* for essential nitrogen in the form of  $NH_{4}^{+}$  (Figure 1b).

Stable coexistence was interrogated by altering starting species ratios and by serially transferring cocultures. N<sub>2</sub>-supplied cocultures converged to a common species composition (~10% E. coli) from a range of starting species ratios spanning 12 orders of magnitude (Figure 1d). Once achieved, this ratio was maintained throughout exponential growth and over multiple transfers ( $\geq$ 30 generations; Supplementary Figure 4). Furthermore, cocultures were readily initiated from single colonies grown on nutrientreplete agar (Supplementary Figure 4), demonstrating that metabolic cooperation was immediate and did not require metabolic priming or coevolution. Thus, in this system metabolic interdependency drives ratio convergence and enforces stable coexistence by means of coupled growth.

Stable coexistence was accompanied by reproducible fermentation product yields over serial trans-(Supplementary Figure 4). The soluble fers fermentation products excreted by E. coli are ethanol, formate, lactate, acetate and succinate (Figure 2a). There was no detectable lactate, acetate, or succinate remaining in N<sub>2</sub>-supplied cocultures, suggesting that these fermentation products were consumed by *R. palustris* whereas formate and ethanol were not (Figure 2a). Indeed, R. palustris only consumed lactate, acetate and succinate when fed supernatants from fermentative E. coli monocultures (Supplementary Figure 5). Hereon, we will collectively refer to lactate, acetate and succinate as consumable organic acids. In addition to soluble products, *E. coli* also produces  $H_2$  and  $CO_2$ . As noted above, *R. palustris* Nx is genetically incapable of  $H_2$  consumption.  $CO_2$  fixation by *R. palustris*  Dose-dependent nutrient toxicity in coculture B LaSarre et al



**Figure 2** Growth and metabolic trends of  $NH_4^+$ -supplied cocultures resemble those of *E. coli* monocultures. Metabolic profiles (**a**), specific growth rates (**b**), final cell densities (**c**) and final pH (**d**) of  $NH_4^+$ -supplied or  $N_2$ -supplied cocultures compared with  $NH_4^+$ -supplied *E. coli* monocultures. All cocultures contained *R. palustris* Nx. Data shown is combined from experiments using cultures inoculated using single colonies, subculturing, and/or 1:1 species ratios (cocultures only). Ec. *E. coli*; Rp, *R. palustris*. Error bars, s.e.m.,  $n \ge 3$ .

is likely negligible in cocultures as Calvin cycle transcript levels are low during N<sub>2</sub> fixation (McKinlay and Harwood, 2010; 2011). Corroborating this expectation, cocultures containing an *R. palustris* Nx Calvin cycle mutant, incapable of  $CO_2$  fixation, behaved comparably to cocultures containing *R. palustris* Nx (Supplementary Figure 6).

To test the importance of metabolic interdependence for coculture stability, we initiated cocultures where we externally supplied either acetate, a carbon source for R. palustris, or  $NH_4^+$ , a nitrogen source that both species can use (Figure 1a). When acetate was externally supplied, coculture growth rate mirrored that of cocultures supplied glucose alone, even over serial transfers (Supplementary Figure 7). In addition, in acetate-supplied cocultures *E. coli* grew to final cell densities comparable to those in N<sub>2</sub>-supplied cocultures with glucose alone (Supplementary Figure 7). Within our coculture system, organic acids serve as both the essential carbon source for *R. palustris* growth as well as the electron source for  $N_2$  fixation. Hence, these data signify that growth and  $NH_4^+$  cross-feeding by R. palustris in N<sub>2</sub>-supplied cocultures are not limited by either carbon or electron availability.

In contrast to acetate, external supply of NH<sub>4</sub><sup>+</sup> drastically altered coculture behavior. When NH<sub>4</sub><sup>+</sup> was supplied, we observed rapid growth regardless of the *R. palustris* genotype (Figures 1c and 2b). NH<sub>4</sub><sup>+-</sup> supplied cocultures were white, unlike N<sub>2</sub>-supplied which were red (Supplementary cocultures. Figure 8), suggesting that NH<sub>4</sub><sup>+</sup>-supplied cocultures were dominated by E. coli rather than R. palustris. E. coli dominance was confirmed by both FISH and selective plate counts (Figures 1d and 2c). Growth, fermentation profiles, and final pH of NH<sup>+</sup><sub>4</sub>-supplied cocultures mirrored those of *E. coli* monocultures (Figure 2), indicating negligible metabolic contribution by *R. palustris* in these cocultures. Specifically, we observed the accumulation consumable organic acids (Figure 2a). The lack of conversion of consumable organic acids into R. palustris biomass in NH<sub>4</sub>supplied cocultures resulted in final cell densities that were much lower than those of N<sub>2</sub>-supplied

cocultures (Figure 1c). Overall, these data demonstrate that the external supply of  $NH_4^+$  breaks the reciprocal dependency of this syntrophy and allows *E. coli* to rapidly overtake the coculture due to its lower intrinsic generation time relative to that of *R. palustris* (2.5 h versus 9.9 h), and ultimately prevent *R. palustris* growth. These results, in combination with the acetate results discussed above, also corroborate the notion that stable coexistence relies on the metabolic dependence of the faster-growing species on its slower-growing partner and not the converse (Hom and Murray, 2014).

## Organic acids have both positive and negative roles within the mutualism

It initially seemed paradoxical that there was negligible R. palustris growth in NH<sub>4</sub><sup>+</sup>-supplied cocultures despite the presence of consumable organic acids (Figure 2). However, we noticed that the final pH of NH<sub>4</sub><sup>+</sup>-supplied cocultures was much lower than that of N<sub>2</sub>-supplied cocultures (Figure 2d). On the basis of this observation, we reasoned that organic acids could have two roles in our system: a positive role as a carbon source for R. palustris and a negative role when abundant through culture acidification. To test this hypothesis, we monitored growth and fermentation product yields in NH<sub>4</sub><sup>+</sup>-supplied cocultures that were additionally buffered with 100 mm MOPS, pH 7. We expected that the additional buffer would keep the pH sufficiently neutral to allow *R. palustris* growth. In agreement with this hypothesis, NH<sub>4</sub><sup>+</sup>-supplied cocultures were indeed able to support *R. palustris* growth when supplemented with MOPS (Figure 3a; Supplementary Figure 8). Under these sufficiently buffered conditions, E. coli still grew rapidly and fermentation products accumulated (Figure 3); however, slower-growing R. palustris eventually assimilated the accumulated consumable organic acids and reached cell densities comparable to those seen in N<sub>2</sub>-supplied cocultures (Figure 3c; compare with Figure 2c). In a complementary approach, we also grew R. palustris in spent E. coli monoculture

342



**Figure 3** Rapid fermentation by *E. coli* in NH<sub>4</sub><sup>+</sup>-supplied cocultures results in a growth-inhibiting pH. (a) Growth curves with corresponding final pH values of NH<sub>4</sub><sup>+</sup>-supplied cocultures containing *R. palustris* Nx grown in MDC without or with additional 100 mM MOPS buffer, pH 7 (-/+ MOPS). Cocultures were inoculated by subculturing. Log transformed growth curves are in Supplementary Figure 3. (b, c) Fermentation product yields (b) and cell densities (c) at the intermediate and final time points indicated in **a**. Error bars, s. e.m., n=3. Some error bars are too small to visualize.

supernatants that were either unmodified (– MOPS; pH 4.9) or were additionally buffered with 100 mm MOPS, pH 7 (+MOPS; pH 6.2) before inoculation with *R. palustris*. No growth was observed in spent supernatants without MOPS. In contrast, in MOPS-buffered *E. coli* supernatants *R. palustris* assimilated consumable organic acids and grew to a similar cell density as that observed in N<sub>2</sub>-supplied cocultures (Supplementary Figure 5). The results from these two experiments corroborated that in NH<sub>4</sub><sup>+</sup>-supplied cocultures without MOPS, the rapid accumulation of organic acids quickly shifts the role of organic acids from being a nutrient to being harmful to *R. palustris*.

## Development of an ecological model capable of predicting coculture behavior

Mathematical modeling of synthetic communities is a powerful tool for generating experimentally-



**Figure 4** Schematic of the ecological model. Ec and Rp represent *E. coli* and *R. palustris.* Orange arrows indicate that production can be growth-independent. Red oval arrows represent inhibitory effects of organic acids; this parameter is included in model A but not in model B. Bracketed letters indicate model code designations (see 'Materials and methods' section, Supplementary Table 2). EtOH, ethanol; For, formate; Glc, glucose; OAcs, consumable organic acids (lactate, acetate and succinate).

testable predictions of what controls mutualism dynamics and stability (Klitgord and Segrè, 2010; Mee *et al.*, 2014; Zomorrodi and Segrè, 2015; Zelezniak *et al.*, 2015). We therefore developed an ecological model (Figure 4) based on previously described interactions for mutualistic systems with Monod uptake kinetics (Meyer *et al.*, 1975; Lee *et al.*, 1976). We modified the model to describe batch culturing and to include growth-independent fermentation by *E. coli* (Supplementary Table 2; Supplementary Figure 1; Wanner and Egli, 1990). Our model simulates NH<sup>4</sup><sub>4</sub> excretion in a growthdependent manner because the essential carbon source for *R. palustris* growth (organic acids) is also the electron source for N<sub>2</sub> fixation.

We also modified our model to describe acid inhibition of both *R. palustris* and *E. coli* growth by organic acids (model A). This modification was necessary to correctly predict resource and population dynamics in NH<sup>+</sup><sub>4</sub>-supplied cocultures, specifically domination of the coculture by *E. coli*, organic acid accumulation and inhibition of *R. palustris* growth (Figure 5b, left panel; compare with Figure 5a left panel and Figure 2). When acid inhibition was omitted from the model (model B), the model predicted growth and metabolic trends in NH<sub>4</sub>supplied cocultures that were inconsistent with our experimental observations. Specifically, model B incorrectly predicted that rapid *E. coli* growth would be accompanied by slower *R. palustris* Nx growth in NH<sup>+</sup><sub>4</sub>-supplied cocultures and that final metabolic profiles and species densities would be comparable to N<sub>2</sub>-supplied cocultures (Figure 5c, left panel; compare with Figure 5c, right panel). These predictions were, however, consistent with the trends observed in MOPS-buffered NH<sup>+</sup><sub>4</sub>-supplied cocultures (compare Figure 5c left panel with Figure 3), further substantiating both the impact of acid inhibition in NH<sup>+</sup><sub>4</sub>-supplied cocultures and the utility of the ecological model in predicting how specific factors influence coculture behavior.

Predictions for  $N_2$ -supplied cocultures were near identical using the two models (Figures 5b and c,



Dose-dependent nutrient toxicity in coculture

**Figure 5** Ecological modeling accurately predicts coculture behavior when growth-inhibiting effects of organic acids are included. Experimental (a) and simulated (b, c) growth and metabolic profiles of *E. coli+R. palustris* Nx NH<sub>4</sub><sup>+</sup>-supplied (left) or N<sub>2</sub>-supplied (right) cocultures. (a) All cocultures were inoculated by subculturing. Most error bars (s.e.m., n = 3) are too small to visualize. (b, c) Simulations were run using model A (b) or model B (c) with default parameters (Supplementary Table 2) except that for simulations with NH<sub>4</sub><sup>+</sup> (left panels) the initial NH<sub>4</sub><sup>+</sup> concentration was set to 15 mM. Ec, *E. coli*; Rp, *R. palustris*; OAcs, consumable organic acids.

right panels). Both models predicted that consumable organic acids would only briefly accumulate, and that the coculture would be primarily composed of *R. palustris* Nx (14% *E. coli* simulated versus ~10% experimental; Figure 5, right panels). These simulations were corroborated by our empirical data showing that additional buffer had no observable effect on growth or metabolic trends in N<sub>2</sub>-supplied cocultures (Supplementary Figure 9). Thus, the inhibitory effects of organic acids are not observed in N<sub>2</sub>-supplied cocultures as consumption by *R. palustris* keeps pace with production. In other words, balanced metabolic interactions in N<sub>2</sub>-supplied cocultures self-buffered the system.

# High cross-feeding levels can be detrimental to a mutualism

Cross-feeding levels are inherently difficult to measure and yet are hypothesized to be a major determinant of mutualism dynamics and stability (Shou *et al.*, 2007; Kim *et al.*, 2008; Estrela *et al.*,

The ISME Journal

344

2012; Hom and Murray, 2014). We therefore used our model to address the effect of NH<sup>+</sup><sub>4</sub>-cross-feeding levels on mutualism dynamics. By varying the NH<sub>4</sub><sup>+</sup> excretion parameter we estimated the level of NH<sub>4</sub><sup>+</sup> excretion by R. palustris Nx that would result in observed coculture trends (Figures 6a, 1×). A large range of NH<sup>4</sup> excretion, both higher and lower than that of *R. palustris* Nx and spanning over 3 orders of magnitude, was predicted to support cooperative growth (Figure 6a). Surprisingly, at levels above that excreted by R. palustris Nx, the model predicted that increasing the amount of NH<sup>+</sup><sub>4</sub> excreted per R. palustris cell would decrease the R. palustris carrying capacity (the maximum population size allowed by the culture conditions; Figure 6a). More  $NH_{4}^{+}$  excretion would result in faster *E. coli* growth and organic acid excretion, outpacing consumption by *R. palustris*. Consequently, organic acids would accumulate, acidify the medium and decrease the R. palustris carrying capacity (Figure 6a). The acid inhibition of *R. palustris* growth predicted to result





а

OAcs For

Ec,

**10**<sup>10</sup>

10<sup>9</sup>

Figure 6 Higher NH<sup>+</sup><sub>4</sub> excretion levels lead to cocultures with less R. palustris and less carbon assimilation. (a) Simulated effect of the R. palustris  $NH_4^+$  excretion level on growth and organic acid accumulation in N<sub>2</sub>-supplied cocultures. Concentration of 1× is the default NH<sub>4</sub><sup>+</sup> excretion level (0.25 fmol NH<sub>4</sub><sup>+</sup>/cell) and is thought to represent that excreted by R. palustris Nx, based on model approximation of experimental trends. Concentration of  $3 \times$ (0.75 fmol NH<sub>4</sub>/cell) indicates an NH<sub>4</sub> excretion level thought to represent that by R. palustris Nx $\Delta$ AmtB (CGA4021), based on monoculture  $NH_4^+$  levels compared with those of *R. palustris* Nx (Supplementary Figure 2). OAcs, consumable organic acids; For, formate; Eci and Rpi, initial E. coli and R. palustris cell densities (dashed lines); Ecf and Rpf, final E. coli and R. palustris cell densities (solid lines). (**b**- $\hat{\mathbf{d}}$ ) Experimental data from *R. palustris* Nx-based and R. palustris NxAAmtB-based cocultures, including growth curves and final pH values (b), final cell densities (c) and final E. coli percentages over serial transfers (d). Log transformed growth curves are in Supplementary Figure 3. All cocultures were inoculated at a 1:1 species ratio. Error bars, s.e.m., n=3.

from higher NH<sup>+</sup><sub>4</sub> excretion is analogous to trends observed when cocultures were supplemented with  $NH_{4}^{+}$ , though the effects from the  $NH_{4}^{+}$  supplement are predicted to be more severe (compare Figure 6a with

These predictions of altered community composition highlighted a crucial aspect of this mutualism: the inhibitory activity of organic acids is directly proportional to their net excretion level. Consequently, the level of cross-feeding determines the equilibrium ratio and carrying capacity for each species. Moreover, contradictory as it may seem, enhanced cross-feeding is predicted to actually be detrimental to this mutualism (lower R. palustris carrying capacity and less carbon assimilated by the community).

To experimentally test these predictions, we engineered R. palustris to excrete more  $NH_4^+$  than the Nx strain. In several diazotroph species, disruption of the NH<sub>4</sub><sup>+</sup> transporter, AmtB, results in extracellular NH<sup>+</sup><sub>4</sub> accumulation during N<sub>2</sub> fixation (Yakunin and Hallenbeck, 2002; Zhang et al., 2012; Barney et al., 2015). We deleted the genes for both AmtB homologs in R. palustris Nx, resulting in *R. palustris* Nx $\Delta$ AmtB, and found that NH<sup>+</sup><sub>4</sub> accumulated to nearly three times to that of *R. palustris* Nx in monocultures (Supplementary Figure 2). The increased  $NH_4^+$  accumulation was due to the *amtB2* deletion; the *amtB1* deletion had no effect (Supplementary Figure 2). When we paired R. palustris  $Nx\Delta AmtB$  with E. coli, coculture behavior matched that predicted by the modeled  $3 \times \text{NH}_{4}^{+}$  excretion level (Figure 6a). The combined carrying capacity was lower than that of cocultures with R. palustris Nx (Figure 6b), R. palustris cell densities declined (Figure 6c), the final species ratio changes to approximately 1:1 (Figure 6c), and consumable organic acids accumulated and acidified the medium (Figure 6b; Supplementary Figure 10). Growth yield and organic acid assimilation levels in cocultures with *R. palustris* Nx $\Delta$ AmtB were restored to levels seen in cocultures with *R. palustris* Nx by adding MOPS buffer, confirming that acidification was prematurely inhibiting R. palustris  $Nx\Delta AmtB$ growth (Supplementary Figure 10).

Despite the detrimental effect of increased NH<sub>4</sub><sup>+</sup> excretion on the R. palustris carrying capacity, metabolic and growth trends were reproducible and stable coexistence was maintained over multiple transfers (Figure 6c; Supplementary Figure 10). Thus, the obligate nature of cross-feeding within this mutualism ensured stable coexistence even at a new equilibrium state. The model predicted that different NH<sub>4</sub><sup>+</sup> excretion levels would establish a range of stable equilibria (Figure 6a). These distinct equilibria are likely all stabilized by negative frequency-dependent selection on E. coli (Morris, 2015), wherein the relative amount of NH<sup>+</sup><sub>4</sub> available per E. coli cell, which determines E. coli fitness, decreases as the *E. coli* frequency increases. The level of NH<sub>4</sub><sup>+</sup> exchange also influenced organic acid

transformation by *R. palustris* into biomass and  $CO_2$  (Figure 6a; Supplementary Figure 10) and thus impacted the nature and efficiency with which this synthetic community transformed carbon.

## Discussion

Organic acids are important drivers of anaerobic food web ecology, acting as a resource for some microbes while inhibiting others (Russell and Diez-Gonzalez, 1997; Schink, 2002; McInerney et al., 2009). In our synthetic community, the organic acids cross-fed by *E. coli* function as both a nutrient and an inhibitor to R. palustris. This attribute crucially impacts how cross-feeding levels govern mutualism dynamics. Specifically, excessive cooperation by *R. palustris* (that is, increased  $NH_4^+$  cross-feeding) stimulates reciprocation by E. coli, resulting in organic acid production that outpaces consumption by *R. palustris*. Consequently, increased cooperation moves the nature of organic acids along a continuum from beneficial to detrimental, limiting total coculture growth and carbon transformation. Presumably, the influence of dose-dependent metabolite toxicity on a mutualism depends on at least three factors: (i) the rate of excretion, which depends on the producer population size; (ii) the rate of consumption or detoxification by the partner, which depends on the consumer population size; and (iii) the metabolite's inhibitory activity. For example, a highmetabolite excretion level and/or a high-metabolite toxicity would more readily impair cooperation. In contrast, enhanced consumption by the partner would facilitate cooperation at higher exchange levels (Supplementary Figure 11). Thus, the detrimental influence exerted by a metabolite on the system will depend on the activity of both partners.

The closed nature of our experimental system circumvents two other factors that could also influence cross-feeding and dose-dependent nutrient toxicity in natural systems: diffusion of metabolites in an open system and metabolic contributions by additional species. Both of these factors would influence the accumulation of a metabolite and therefore determine its position on the continuum between beneficial nutrient and detrimental toxin. Synthetic ecosystems such as the one described here could potentially be modified to understand how such additional factors influence dose-dependent nutrient toxicity and thereby mutualism behavior.

Separately, we have also shown herein that  $NH_4^+$  cross-feeding by *R. palustris* can result in an immediate and stable syntrophic relationship with fermentative *E. coli* under conditions requiring  $N_2$  fixation. Metabolite excretion is a mechanism by which cooperative interactions can be initiated and sustained (Sachs *et al.*, 2004). There is accumulating evidence that some diazotrophs excrete  $NH_4^+$  during  $N_2$  fixation (Adam *et al.*, 2016) and that AmtB functions to reacquire  $NH_4^+$  and limit its availability

to nearby organisms (Supplementary Figure 2) (Yakunin and Hallenbeck, 2002; Zhang et al., 2012; Barney et al., 2015). As proposed in the Black Queen hypothesis (Morris et al., 2012), this NH<sup>+</sup><sub>4</sub> leakage makes diazotrophs well-suited to establish nascent mutualisms, both in natural and synthetic communities. Our results support this notion. Furthermore, the specific ability of enhanced NH<sup>+</sup><sub>4</sub> cross-feeding to elevate reciprocal exchange of potentially toxic compounds could have implications for the deployment of engineered NH<sup>+</sup><sub>4</sub>-excreting diazotrophs as biofertilizers for agricultural and industrial applications (Ortiz-Marquez et al., 2012; Barney et al., 2015; Geddes et al., 2015): Increasing NH<sup>+</sup><sub>4</sub> excretion might not equate to optimizing  $NH_4^+$  excretion. Overall, our results indicate that variations in cross-feeding levels have the potential to alter the nature of interspecies relationships and potentially ecosystem function.

## **Conflict of Interest**

The authors declare no conflict of interest.

## Acknowledgements

We thank A Posto for early contributions; Prof SR Hall for modeling discussions; Prof Y Brun for critical reading; and RK Fritts and the McKinlay lab for discussions. This work was supported by the US Department of Energy, Office of Science, Office of Biological and Environmental Research, under Award Number DE-SC0008131. Model development was supported in part by the US Army Research Office, Grant W911NF-14-1-0411.

### References

- Abbott BJ. (1973). Ethanol inhibition of a bacterium (*Acinetobacter calcoaceticus*) in chemostat culture. J Gen Microbiol **75**: 383–389.
- Adam B, Klawonn I, Svedén JB, Bergkvist J, Nahar N, Walve J *et al.* (2016). N<sub>2</sub>-fixation, ammonium release and N-transfer to the microbial and classical food web within a plankton community. *ISME J* **10**: 450–459.
- Alvarez LA, Exton DA, Timmis KN, Suggett DJ, McGenity TJ. (2009). Characterization of marine isoprene-degrading communities. *Environ Microbiol* **11**: 3280–3291.
- Barney BM, Eberhart LJ, Ohlert JM, Knutson CM, Plunkett MH. (2015). Gene deletions resulting in increased nitrogen release by *Azotobacter vinelandii*: Application of a novel nitrogen biosensor. *Appl Environ Microbiol* **81**: 4316–4328.
- Barnhill AE, Weeks KE, Xiong N, Day TA, Carlson SA. (2010). Identification of multiresistant Salmonella isolates capable of subsisting on antibiotics. Appl Environ Microbiol 76: 2678–2680.
- Ding J, Liu BF, Ren NQ, Xing DF, Guo WQ, Xu JF *et al.* (2009). Hydrogen production from glucose by co-culture of *Clostridium butyricum* and immobilized

Rhodopseudomonas faecalis RLD-53. Int J Hydrogen Energy **34**: 3647–3652.

- Ducret A, Quardokus E, Brun Y. (2016). MicrobeJ, a high throughput tool for quantitative bacterial cell detection and analysis. *Nat Microbiol* 1: 16077.
- Estrela S, Trisos CH, Brown SP. (2012). From metabolism to ecology: cross-feeding interactions shape the balance between polymicrobial conflict and mutualism. *Am Nat* **180**: 566–576.
- Fang HHP, Zhu H, Zhang T. (2006). Phototrophic hydrogen production from glucose by pure and co-cultures of *Clostridium butyricum* and *Rhodobacter sphaeroides*. *Int J Hydrogen Energy* **31**: 2223–2230.
- Fuhrman JA. (2009). Microbial community structure and its functional implications. *Nature* **459**: 193–199.
- Geddes BA, Ryu M-H, Mus F, Garcia Costas A, Peters JW, Voigt CA *et al.* (2015). Use of plant colonizing bacteria as chassis for transfer of  $N_2$ -fixation to cereals. *Curr Opin Biotechnol* **32**: 216–222.
- Harcombe W. (2010). Novel cooperation experimentally evolved between species. *Evolution* **64**: 2166–2172.
- Hillesland KL, Stahl DA. (2010). Rapid evolution of stability and productivity at the origin of a microbial mutualism. *Proc Natl Acad Sci USA* **107**: 2124–2129.
- Hom EFY, Murray AW. (2014). Niche engineering demonstrates a latent capacity for fungal-algal mutualism. *Science* 345: 94–88.
- Huang CB, Alimova Y, Myers TM, Ebersole JL. (2011). Short- and medium-chain fatty acids exhibit antimicrobial activity for oral microorganisms. *Arch Oral Biol* 56: 650–654.
- Huang JJ, Heiniger EK, McKinlay JB, Harwood CS. (2010). Production of hydrogen gas from light and the inorganic electron donor thiosulfate by *Rhodop*seudomonas palustris. Appl Environ Microbiol **76**: 7717–7722.
- Jiao Y, Navid A, Stewart BJ, McKinlay JB, Thelen MP, Pett-Ridge J. (2012). Syntrophic metabolism of a co-culture containing *Clostridium cellulolyticum* and *Rhodop*seudomonas palustris for hydrogen production. Int J Hydrogen Energy 37: 11719–11726.
- Johns NI, Blazejewski T, Gomes AL, Wang HH. (2016). Principles for designing synthetic microbial communities. Curr Opin Microbiol 31: 146–153.
- Kim HJ, Boedicker JQ, Choi JW, Ismagilov RF. (2008). Defined spatial structure stabilizes a synthetic multispecies bacterial community. *Proc Natl Acad Sci USA* 105: 18188–18193.
- Kim M, Harwood CS. (1991). Regulation of benzoate-CoA ligase in *Rhodopseudomonas palustris*. *FEMS Micrbiol Lett* 83: 199–203.
- Klitgord N, Segrè D. (2010). Environments that induce synthetic microbial ecosystems. *PLoS Comput Biol* 6: e1001002.
- Kremer TA, LaSarre B, Posto AL, McKinlay JB. (2015).  $N_2$  gas is an effective fertilizer for bioethanol production by *Zymomonas mobilis*. *Proc Natl Acad Sci USA* **112**: 2222–2226.
- Kunz DA, Nagappan O, Silva-Avalos J, Delong GT. (1992). Utilization of cyanide as nitrogenous substrate by *Pseudomonas fluorescens* NCIMB 11764: evidence for multiple pathways of metabolic conversion. *Appl Environ Microbiol* 58: 2022–2029.
- Larimer FW, Chain P, Hauser L, Lamerdin J, Malfatti S, Do L et al. (2004). Complete genome sequence of the metabolically versatile photosynthetic bacterium *Rho*dopseudomonas palustris. Nat Biotechnol 22: 55–61.

- Lee IH, Fredrickson AG, Tsuchiya HM. (1976). Dynamics of mixed cultures of *Lactobacillus plantarum* and *Propionibacterium shermanii*. *Biotechnol Bioeng* **18**: 513–526.
- Lindemann SR, Bernstein HC, Song H-S, Fredrickson JK, Fields MW, Shou W *et al.* (2016). Engineering microbial consortia for controllable outputs. *ISME J* **10**: 2077–2084.
- Louis P, Scott KP, Duncan SH, Flint HJ. (2007). Understanding the effects of diet on bacterial metabolism in the large intestine. J Appl Microbiol **102**: 1197–1208.
- McInerney MJ, Sieber JR, Gunsalus RP. (2009). Syntrophy in anaerobic global carbon cycles. *Curr Opin Biotechnol* **20**: 623–632.
- McKinlay JB, Harwood CS. (2011). Calvin cycle flux, pathway constraints, and substrate oxidation state together determine the  $H_2$  biofuel yield in photoheter-otrophic bacteria. *mBio* **2**: e00323–10.
- McKinlay JB, Harwood CS. (2010). Carbon dioxide fixation as a central redox cofactor recycling mechanism in bacteria. *Proc Natl Acad Sci USA* **107**: 11669–11675.
- McKinlay JB, Zeikus JG, Vieille C. (2005). Insights into *Actinobacillus succinogenes* fermentative metabolism in a chemically defined growth medium. *Appl Environ Microbiol* **71**: 6651–6656.
- Mee MT, Collins JJ, Church GM, Wang HH. (2014). Syntrophic exchange in synthetic microbial communities. *Proc Natl Acad Sci USA* **111**: E2149–E2156.
- Meyer JS, Tsuchiya HM, Fredrickson AG. (1975). Dynamics of mixed populations having complementary metabolism. *Biotechnol Bioeng* **17**: 1065–1081.
- Momeni B, Chen C-C, Hillesland KL, Waite A, Shou W. (2011). Using artificial systems to explore the ecology and evolution of symbioses. *Cell Mol Life Sci* 68: 1353–1368.
- Morris BEL, Henneberger R, Huber H, Moissl-Eichinger C. (2013). Microbial syntrophy: interaction for the common good. *FEMS Microbiol Rev* **37**: 384–406.
- Morris JJ. (2015). Black Queen evolution: the role of leakiness in structuring microbial communities. *Trends Genet* **31**: 475–482.
- Morris JJ, Lenski RE, Zinser ER. (2012). The Black Queen Hypothesis: evolution of dependencies through adaptive gene loss. *mBio* **3**: e00036–12.
- Oda Y, Šlagman S-J, Meijer WG, Forney LJ, Gottschal JC. (2000). Influence of growth rate and starvation on fluorescent *in situ* hybridization of *Rhodopseudomonas palustris. FEMS Microbiol Ecol* **32**: 205–213.
- Odom JM, Wall JD. (1983). Photoproduction of  $H_2$  from cellulose by an anaerobic bacterial coculture. Appl Environ Microbiol **45**: 1300–1305.
- Ortiz-Marquez JCF, Do Nascimento M, Dublan M de LA, Curatti L. (2012). Association with an ammoniumexcreting bacterium allows diazotrophic culture of oilrich eukaryotic microalgae. *Appl Environ Microbiol* **78**: 2345–2352.
- Pelletier DA, Hurst GB, Foote LJ, Lankford PK, McKeown CK, Lu TY *et al.* (2008). A general system for studying protein-protein interactions in Gram-negative bacteria. *J Proteome Res* **7**: 3319–3328.
- Rey FE, Oda Y, Harwood CS. (2006). Regulation of uptake hydrogenase and effects of hydrogen utilization on gene expression in *Rhodopseudomonas palustris*. *J Bacteriol* **188**: 6143–6152.

- Russell JB, Diez-Gonzalez F. (1997). The Effects of fermentation acids on bacterial growth. *Adv Microb Physiol* **39**: 205–234.
- Sachs JL, Mueller UG, Wilcox TP, Bull JJ. (2004). The evolution of cooperation. *Q Rev Biol* **79**: 135–160.
- Schink B. (2002). Synergistic interactions in the microbial world. Antonie Van Leeuwenhoek 81: 257–261.
- Shou W, Ram S, Vilar JMG. (2007). Synthetic cooperation in engineered yeast populations. Proc Natl Acad Sci USA 104: 1877–1882.
- Sousa DZ, Smidt H, Alves MM, Stams AJM. (2009). Ecophysiology of syntrophic communities that degrade saturated and unsaturated long-chain fatty acids. *FEMS Microbiol Ecol* **68**: 257–272.
- Summers ZM, Fogarty HE, Leang C, Franks AE, Malvankar NS, Lovley DR. (2010). Direct exchange of electrons within aggregates of an evolved syntrophic coculture of anaerobic bacteria. *Science* **330**: 1413–1415.
- Sun Q, Xiao W, Xi D, Shi JP, Yan X, Zhou ZH. (2010). Statistical optimization of biohydrogen production from sucrose by a co-culture of *Clostridium acidisoli* and *Rhodobacter sphaeroides*. Int J Hydrogen Energy 35: 4076–4084.

- Wanner U, Egli T. (1990). Dynamics of microbial growth and cell composition in batch culture. *FEMS Microbiol Lett* **75**: 19–43.
- Widder S, Allen RJ, Pfeiffer T, Curtis TP, Wiuf C, Sloan WT et al. (2016). Challenges in microbial ecology: building predictive understanding of community function and dynamics. *ISME J* 10.1038/ismej.2016.45.
- Yakunin AF, Hallenbeck PC. (2002). AmtB is necessary for NH<sup>+</sup><sub>4</sub>-induced nitrogenase switch-off and ADPribosylation in *Rhodobacter capsulatus*. J Bacteriol **184**: 4081–4088.
- Zelezniak A, Andrejev S, Ponomarova O, Mende DR, Bork P, Patil KR. (2015). Metabolic dependencies drive species co-occurrence in diverse microbial communities. *Proc Natl Acad Sci USA* **112**: 6449–6454.
- Zhang T, Yan Y, He S, Ping S, Alam KM, Han Y *et al.* (2012). Involvement of the ammonium transporter AmtB in nitrogenase regulation and ammonium excretion in *Pseudomonas stutzeri* A1501. *Res Microbiol* **163**: 332–339.
- Zomorrodi AR, Segrè D. (2015). Synthetic ecology of microbes: mathematical models and applications. *J Mol Biol* **428**: 837–861.

Supplementary Information accompanies this paper on The ISME Journal website (http://www.nature.com/ismej)